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Pulsed-light reduces the toxicity of the algal toxin okadaic acid to freshwater crustacean *Daphnia pulex*

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Abstract

This constitutes the first study to report on the reduction in toxicity of the dinoflagellate algal toxin okadaic acid after novel pulsed light (PL) treatments where ecotoxicological assessment was performed using a miniaturised format of the conventional *in vivo* freshwater crustacean *Daphnia sp.* Acute toxicity test. Bivalves accumulate this toxin, which can then enter the human food chain causing deleterious health effects such as Diarrhetic Shellfish Poisoning. This miniaturised toxicological bioassay used substantially less sample volume and chemical reagents. Findings revealed a 24 h EC₅₀ of 25.87 µg/L for PL-treated okadaic acid at UV dose of 12.98 µJ/cm² compared to a 24 h EC₅₀ of 1.68 µg/L for the untreated okadaic acid control, suggesting a 15-fold reduction in toxicity to *Daphnia pulex*. The bioassay was validated in this study and correlated well with the "classic" ISO format (r = 0.98) using the traditional reference chemical potassium dichromate (K₂Cr₂O₇). Reduction by up to 65% in PL-treated okadaic acid concentration was confirmed by LC-MS/MS analysis. Findings from this study have positive ecological, societal and enterprise implications, such as the development of PL technology for the prevention or reduce algal contamination of fisheries and aquaculture industries.

Keywords

Pulsed-light, Harmful algal blooms, Okadaic acid, Diarrhetic shellfish poisoning, *Daphnia pulex*, Ecotoxicology, Detoxification.

Introduction

Toxic algal blooms are problematic in Ireland and globally. Algal species that produce toxins include *Karenia mikimotoi*, *Microcystis aeruginosa*, *Dinophysis* spp. and *Alexandrium* spp. In the summer of 2012, the Irish Oyster Industry was devastated by toxic blooms of *Karenia mikimotoi*, and in October 2016, the Irish Marine Institutes Weekly Harmful Algal Bloom Bulletin (9–15 October 2016), reported that up to 11% of Irish Aquaculture sites were affected with toxic levels of DSP from *Dinophysis* spp. The Irish Fisheries Board (Bord Iascaigh Mhara 2015) suggest that Irish aquaculture industries were worth an estimated €115 million to the economy in 2014. As most of the aquaculture industry is located in estuarine waters it is susceptible to upstream municipal or agricultural pollution leading to eutrophication. The economic importance is evident as failures or closures of these industries can have an economically devastating effect on communities dependant on aquaculture. The closures of bivalve fisheries off the west coast of Ireland are frequent due to the presence of Diarrhetic Shellfish Poisoning (DSP) causing *Dinophysis* spp. algal blooms and Azaspiracid Poisoning (AZP) in the blue edible mussel *Mytilus edulis*. First reported in 1995, AZP was associated with ingestion of contaminated shellfish from Killary Harbour (Ireland) and has been subsequently implicated in several food poisoning incidents (Twiner *et al.*, 2008) although the symptoms were typical of DSP, the levels of DSP toxins in these Irish shellfish were well below regulatory levels. Consumption of shellfish contaminated with DSP toxins from *Dinophysis* spp., can lead to DSP (Kumagai *et al.*, 1986). DSP toxins include Okadaic Acid (OA) and its derivatives (dinophysistoxins DTX-1 and DTX-3) belong to the lipophilic toxins group (Ben-Gigirey *et al.*, 2007; Luppis *et al.*, 2011). DSP symptoms include diarrhoea, nausea, vomiting and abdominal pain which can appear from 30 min to a few hours after ingestion with complete recovery occurring after several days (Yasumoto *et al.*, 1978).

Thus far, the only means of managing marine biotoxin episodes is exclusion from the food chain by means of monitoring programs (Luppis *et al.*, 2011). Appropriate development of shellfish detoxification approaches would be desirable. Such methods should be rapid, efficient, easy-to-apply and not alter the quality of sensory properties of shellfish. Albeit very limited, methods investigated so far for this purpose include thermal processing, freezing, evisceration, supercritical CO₂ with acetic acid, γ -irradiation and ozonation (FAO, 2016; Luppis *et al.*, 2011). Of these aforementioned interventions, only ozonation resulted in effective toxin reduction (in the range of 6-100%, 25-83% and 21-66% for free OA, OA esters and total OA respectively). Luppis *et al.*, (2011) reported that reduction of OA content was substantially higher in homogenised O₃-treated mussel tissue compared with whole shucked mussels. The appearance and texture of γ -irradiated mussels deteriorated pointing to low potential for commercial use.

There are a number of existing protocols for the measurement of OA which must be maintained at levels below 160 μ g/kg of edible material from live bivalve molluscs as set out in Chapter V (2) of Section VII of Annex III to Regulation (EC) No 853/2004 as amended by (EU) 786/2013, (Food Safety Authority of Ireland, 2015). Annex III of Regulation (EC) 2074/2005 sets out the analytical methods to be used by competent authorities to check compliance with the above named legislations prescribed limits. Analytical methods were reviewed by Prego-Faraldo, *et al.*, (2013) and include both biological and chemical methods. The most widely used and accepted methods are the Mouse BioAssay (MBA) and liquid chromatography. Biochemical methods such as the PP2A inhibition assay and immunological methods are also becoming popular. Chemical methods are usually High Performance Liquid

Chromatography (HPLC) coupled with detection systems such as Mass Spectrometry (LC-MS) or Ultra Violet Detection. In terms of biological assessments, the MBA is most widely accepted and is prescribed under EU Regulation 2074. However, with the associated costs and ethical issues surrounding mammalian test systems, there is a trend towards alternative test systems such as crustacean bioassays. Prego-Faraldo *et al.*, (2013) reports that *Daphnia magna* bioassay lacks sensitivity and cannot completely replace the MBA, although it is inexpensive and capable of measuring Okadaic acid levels tenfold lower than the lower limit of detection of the MBA. This is supported by the findings of Vernoux *et al.*, (1993), who assessed toxicity of OA from mussel extracts to *Daphnia magna* with comparisons to MBA and HPLC methods and reports an EC₅₀ of $15 \pm 1.8 \mu\text{g/L}$ for the *Daphnia magna* bioassay and 85% correlation with HPLC methods. In the last decade substantive efforts focused on the development of alternative analytical tools to the accredited AOAC mouse bioassay of Yasumoto *et al.*, (1978) for okadaic acid detection. The enhancements in LC-MS/MS technology in this time and the provision of analytical standards enabled researchers to develop and validate methods for regulatory analysis (Fux *et al.* 2009, 2011; Gerssen *et al.* 2011). Commission Regulation Number 15/2011 states that the liquid chromatography-mass spectrometric (LC-MS/MS) method should now be applied as the reference method for the measurement of marine lipophilic toxins in shellfish including OA. The implementation of LC-MS/MS allowed the phasing out of the live animal assay in EU Member States and since 31 December 2014 as the standard reference method applied to the regulatory monitoring of OA and other lipophilic toxins in shellfish. For regulatory laboratories the LC-MS/MS method is fit-for-purpose for known regulated toxins though requires expensive high specification equipment, trained personnel for its operation and analytical standards.

The freshwater crustacean *Daphnia magna* has been widely used and validated as an ecotoxicological assessment tool for decades (Vernoux *et al.*, 1993; Prego-Faraldo *et al.*, 2013). The test has been accredited as an ISO test (ISO 6341:2012) and an OECD test (OECD 202 : 2004).

As stated in OECD 202 adopted in April 2004: Guidelines for Testing of Chemicals: *Daphnia sp.*, Acute Immobilisation Test states that *Daphnia magna* Straus is the preferred test species although other suitable *Daphnia* species can be used for example *Daphnia pulex* (OECD 202: 2004). Lilius *et al.*, (2005) also reported no significant difference between *Daphnia magna* and *Daphnia pulex* in a selection of 30 different test chemicals. Adaptations of this assay have been reported by Baumann *et al.*, (2014) for the purposes of testing nanomaterials however the miniaturisation performed in this study further reduced test volumes to 5 organisms / 2.5mL test volume well across four replicate wells. Testing 20 organisms in 4 wells with a total test volume requirement of 10mL maintained the 5% significance level of one immobile daphnid, but further reduced the test material requirements to 25% of the classic glass test.

Pulsed light (PL) technology has received considerable attention as a promising next-generation energy-efficient approach for decontaminating food, packaging and air (Hayes *et al.*, 2013; Garvey *et al.*, 2016; Rowan *et al.*, 2016). This approach kills microorganisms using ultrashort duration pulses of an intense broadband emission spectrum that is rich in UV-C germicidal light (200-280 nm). PL is produced using techniques that multiply power manifold by storing electrical energy in a capacitor over relatively long times (fraction of a second) and releasing it in a short time (millionths or thousands of a second) using sophisticated pulse compression techniques (Rowan *et al.*, 2016). A strong advantage of using pulsed xenon lamps over continuous low to medium pressure conventional UV lamps is that the former has a peak power dissipation, which allows for more microbial inactivation due to greater intensity of the light pulses. Despite significant interest in the development of PL as an alternative or complementary means of disinfection, published studies to date have focused on microbial pathogens including bacteria (Garvey *et al.*, 2016; Rowan *et al.*, 2016), yeast (Farrell *et al.*, 2011), fungi (Rowan, 2011), parasites

(Garvey *et al.*, 2010; Garvey *et al.*, 2014), and viruses (Rowan *et al.*, 2016) for food and water applications. To date, there has been no study on the potential of PL to reduce the toxicity of algal toxins. Therefore, the objective of this timely study was to investigate efficacy of pulsed light as an ‘environmentally-friendly’ decontamination technology for novel reduction or removal of the algal toxin okadaic acid (OA) in water, and to ascertain the ecotoxicological status of PL-treated OA using a modified version of the *Daphnia* assay.

Materials and Methods

***Daphnia pulex* culture:** Laboratory stocks of *Daphnia pulex* were cultured in accordance with ISO6341: 2012. Culturing was carried out in 2L glass beakers in a Phytoincubator @ 20°C ±1°C on 16:8 hr light: dark cycle at 4000 lux in Elendt M4 media and fed fresh *Pseudokirchneriella subcapitata* algae.

Acute toxicity testing: Dilution media was prepared in accordance with ISO 6341: 2012 by dissolving 11.76g, 4.93g, 2.59g and 0.23g of CaCl₂·2H₂O, MgSO₄·7H₂O, NaHCO₃ and KCl (Sigma Aldrich) respectively in 1L of distilled deionized water (dd.H₂O). 125mL of each component was then mixed well to 5L dd.H₂O and aerated with a standard aquarium air pump and air stone for 1 hour. Standard volume (10mL) tests were carried out in accordance with ISO 6341: 2012. Final concentrations tested were 0.0, 0.1, 0.5, 1.0, 1.5, 2.0, 2.5, 5.0 mg/L K₂Cr₂O₇ Reference Chemical (Sigma-Aldrich, Product No. 207802, CAS No. 7778-50-9).

Adapted (Miniaturised) format OA acute toxicity test: Gravid females were isolated from lab stock cultures into small groups in 8.5cm petri dishes prior to hatching. Once hatched, five age synchronised neonates <24hrs old were removed in 2mL culture medium for 0.00 – 50µg/L OA and 1.5mL for 100µg/L OA (to allow for subsequent dilution) to each of four wells of a standard 24 well plate (Sarstedt™) giving a total of 20 daphnids per treatment.

From a working stock of 250 µg/L Okadaic acid ammonium salt from *Prorocentrum concavum* (Sigma-Aldrich, Product No. O8010, CAS No. 155716-06-6) final concentrations tested were 0.00, 0.01, 0.05, 1.00, 5.00, 10.0, 25.0, 50.0 and 100µg/L. Test chemicals were initially made up to 5X required concentrations for 0.00-50µg/L and 2.5X for 100µg/L. From this stock, 0.5mL of each concentration (0.00-50µg/L) and 1.0mL for 100µg/L were pipetted gently into each well as required to achieve a final 1X desired concentration. All dilutions were made in culture medium as described above. Immobilised *Daphnia pulex* were enumerated visually after 24 hours’ exposure. For all tests, any daphnid that failed to mobilise within 15 seconds of gentle agitation with a glass rod was considered to be immobile.

Treatment of OA with PL: A portable pulsed power source (PUV-1, Samtech Ltd., Glasgow) was used to power a low-pressure (60kPa) xenon-filled flashlamp (Heraeus Noblelight XAP type NL4006 series constructed from a clear UV transparent quartz tube) that produced a high-intensity diverging beam of polychromatic pulsed light (as per Naughton *et al.* 2017). This irradiation system produces ultra-short duration pulses of an intense broadband emission. Pulsed light (PL) is produced by storing electricity in a capacitor over relatively long times and releasing it as a short duration pulse using sophisticated pulse compression techniques. The PL has a broadband emission spectrum extending from the UV to the infrared with a rich UV content, and its intensity also depends on the level of voltage applied. The light source has an automatic frequency control function which allows it to operate at 1 pulse per second (pps); this setting was used throughout the study. OA in each respective concentration was treated with PL at increasing fluences (µJ/cm²) in 10 mL static volumes that were dispensed into sterile Petri dishes. Light exposure was homogenous as the xenon lamp measuring 9.75 cm was longer than the 8.5 cm standard diameter Petri dish used to treat the solutions.

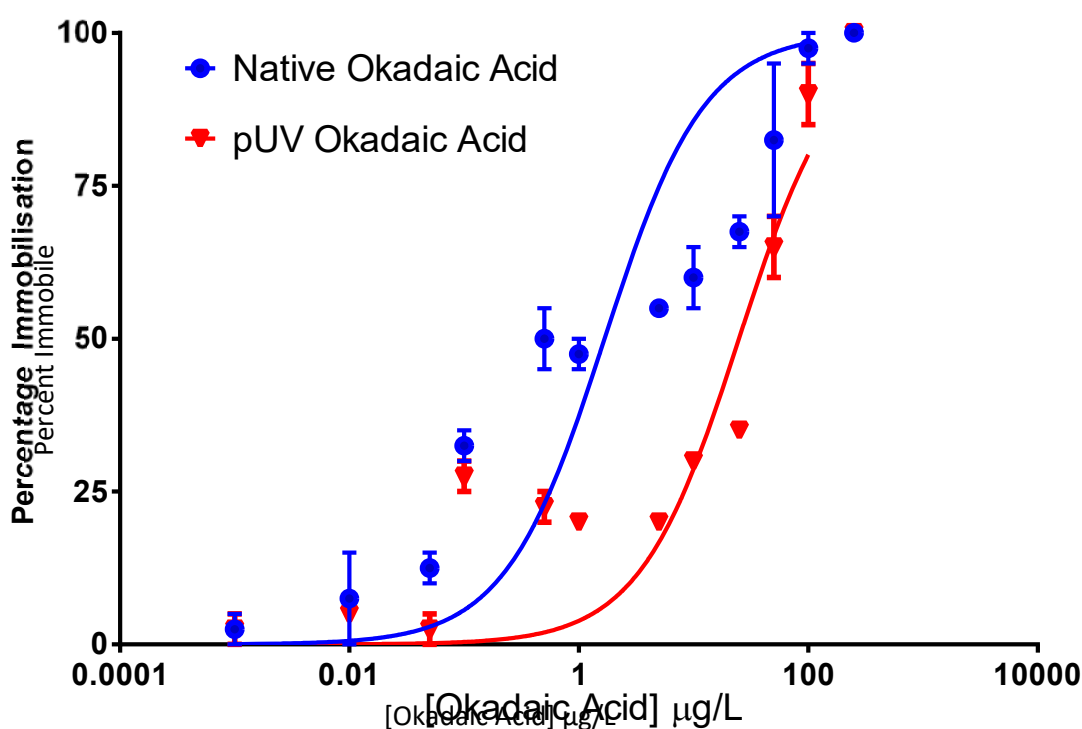
Quantification of test chemicals: Quantification of potassium dichromate was carried out by UV-Vis wavelength scan using a double beam instrument Cary Eclipse™. The scan was performed over the wavelengths 300-650nm. The magnitude of λ_{max} peak of approximately 350nm was used to quantify the samples relative to each other and a standard curve produced.

Two sample sets were analysed, clean and spent. Clean samples are samples of media taken prior to use in the test, i.e. pre-daphnid exposure (0-250µg/L). Spent medium were taken after the test, i.e. post – daphnid exposure (0-100 µg/L). For all samples, (2.5mL replicates) were pooled into 10mL samples and analysed for OA by LC-MS/MS using an Acquity UPLC system coupled to a Quattro premier XE mass spectrometer (Waters, Ireland). The method applied was a modification of the standard operating procedure used by the European Union Reference laboratory for marine biotoxins (EU-RL-MB, Version 5, 2015). Detection and quantification was achieved using targeted analysis via Multiple Reaction Monitoring (MRM) involving the fragmentation of specific precursor ion (parent) using argon as the collision gas, to at least 2 product ions (daughters). The system was operated in electrospray negative mode (ESI⁻) and the parameters were as follows: capillary voltage 2.8 kV, source temperature 120°C, desolvation temperature 400°C, desolvation gas flow 750 L/h. Chromatographic separation was achieved on an acquity UPLC BEH C18 column (100 mm x 2.1 mm id, 1.7 µm particle size) at a column temperature of 45 °C and flow rate of 0.4 ml/min. Mobile phase A was 100 % water and mobile phase B was 90 % acetonitrile/10 % water (v/v) both containing 0.05 % ammonia. A gradient from 20 % to 55% B was run over 5 min, held for 1 min followed by the resetting to initial composition and 2.5 min re-equilibration time. All separations occurred in negative mode. Calibration curves of OA were prepared from 3.125 - 200 ng/ml in clean and spent solutions using GraphPad Prism v6. Retention times of 3.51min were observed following a slight shift from 3.69 min. Transitions used for quantification for OA were 803.35 > 255.1 and 803.35 > 112.9. Samples were also analysed for DTX1 based on the transitions 817.4 > 255.0 and 817.4 > 112.9. However, no DTX1 was observed.

Statistical analysis of ecotoxicological responses: Statistical analysis using GraphPad Prism v6 was carried out to compare all ecotoxicological responses with respective controls. The initial statistical test employed was the Anderson-Darling test for normality. This test found that there was insufficient evidence to classify the data as being from a Gaussian distribution, and therefore the performance of non-parametric tests was suggested. ISO 6341:2012 refers to Gaussian distribution for statistical significance determination. The K₂Cr₂O₇ results were analysed by both parametric and non-parametric tests. These included substitution of the paired t-test with its non-parametric counterpart, the Mann-Whitney U test, and substitution of the Dunnett's test with the non-parametric alternative the Kruskal-Wallis test. Upon comparison of a data from the K₂Cr₂O₇ Classic test in both parametric and non-parametric test suites, the resultant significance values, NOEC (No Observed Effect Concentration), LOEC (Lowest Observed Effect Concentration), MOEC (Maximum Observed Effect Concentration) and EC₅₀ (Effective Median Concentration), were the same irrespective of the parametric strategy. Taking this into account, and given the tendency for other researchers to assume normal distribution and the reference in ISO 6341:2012 to Gaussian distribution, all further tests carried out were the parametric t-test and Dunnett's test and these were used to determine the NOEC and LOEC values on the assumption that additional replication would provide evidence of normality. Pearson Correlation was used to derive a correlation coefficient for the adapted method.

Results

Untreated OA was found to be 15 fold more toxic to *Daphnia pulex* than PL treated OA. The EC_{50} for untreated OA was $1.68\mu\text{g/L}$ (95%CI $0.79\text{--}3.55\mu\text{g/L}$) whilst the EC_{50} for PL treated OA was observed to be $25.87\mu\text{g/L}$ (95%CI $16.68\text{--}40.10\mu\text{g/L}$) as shown in Fig 1. A concomitant reduction in detected OA was also observed by LC-MS/MS following PL treatment. Figure 2 provides graphical data on the reduction in the detectable OA by LC-MS/MS following PL treatment. Nominal concentrations of 0.1 and $0.5\mu\text{g/L}$ were both below the limit of detection post PL treatment and the maximal concentration of detected OA $100\mu\text{g/L}$ was reduced by over 65% post PL treatment. No significant difference was observed in EC_{50} of OA in the 680 nm test (95% CI: 0.7942 to 3.552 digital ISO 6341: 2012 test format and the digital UV-Okadaic Acid test $\mu\text{g/L}$ (95% CI: 16.68 to 40.10 digital ISO 6341:2012, see Table 1. The test formats were correlated using Pearson Correlation. The Correlation Coefficient was 0.9883 (95% CI $0.9334 - 0.9980$) as shown in Fig 3 below.



Dose Response Curve for Okadaic Acid *Daphnia pulex*

Fig 1: Concentration effect curves of okadaic acid in native and PL treated forms illustrating toxicity to *Daphnia pulex* over 24 hours.

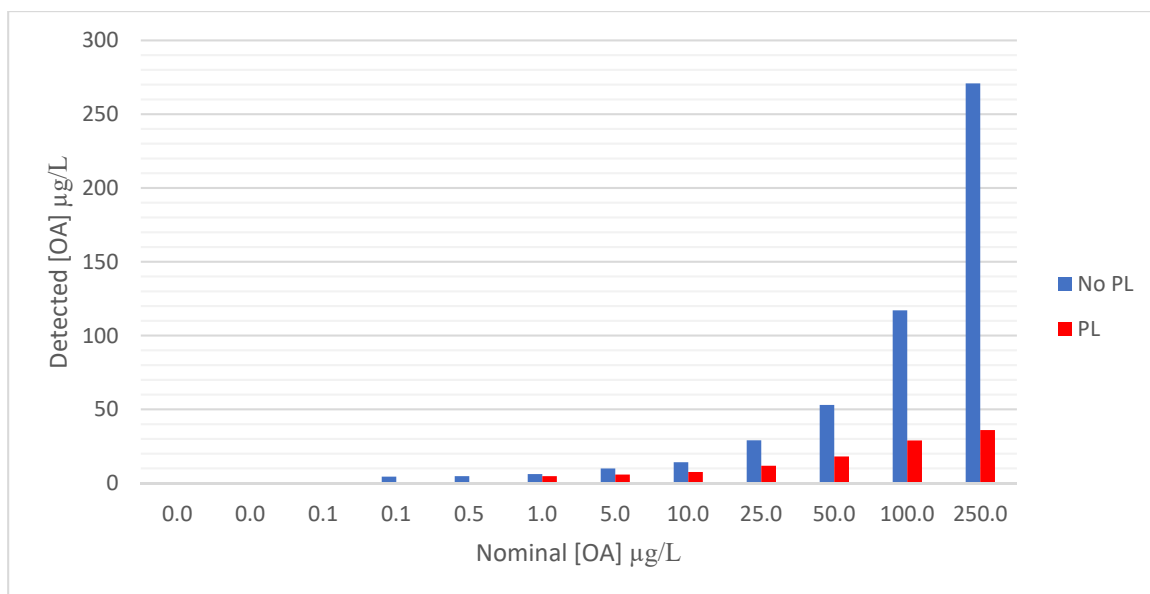


Fig 2: Comparison of okadaic acid nominal concentration with LC-MS/MS detected okadaic acid in native and PL treated forms from spent medium samples.

Table 1: Toxicity assessment of okadaic acid and PL treated OA demonstrating toxicity to *Daphnia pulex* over 24 hours.

Attribute	Okadaic Acid	PL Treated Okadaic Acid
EC ₅₀	1.68µg/L (0.79 – 3.55 g/L) ^a	25.87µg/L (16.68 – 40.10µg/L) ^a
NOEC	0.05µg/L	0.05µg/L
LOEC	0.10µg/L	1.00µg/L
MOEC	100.00 µg/L	100.00 µg/L

a. 95% Confidence Interval

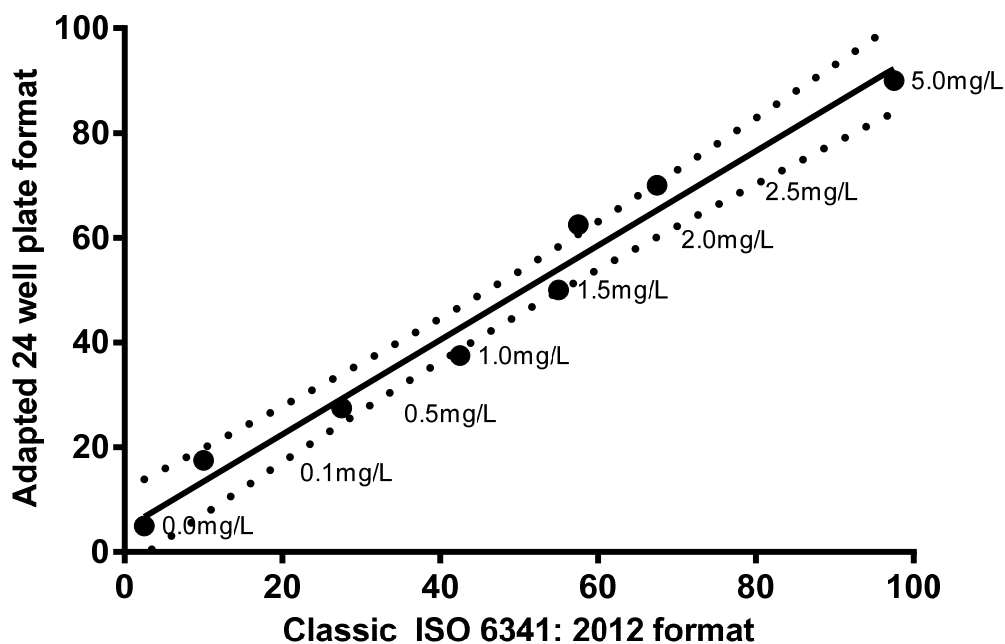


Fig 3: Correlation of miniaturised 24 hr method with classic 24hr ISO method. Pearson Correlation Coefficient 0.9883 (95% Confidence Interval 0.9344 – 0.9980)

Table 2 below provides the percentage loss of OA detection by LC-MS/MS pre and post PL treatment. Figures 4 and 5 provide LC-MS/MS chromatograms of the 100µg/L spent samples also pre and post PL treatment. Supplementary documentation accompanying this paper provide all chromatograms and raw data.

Table 2: Comparison of detected OA by LC-MS/MS pre and post PL treatment

Nominal OA [µg/L]	Average OA [µg/L] untreated	Average OA [µg/L] PL treated	% drop following PL treatment
0	ND	ND	ND
0.01	ND	ND	ND
0.05	ND	ND	ND
0.1	4.45	ND	ND
0.5	4.8	ND	ND
1.0	6.2	4.75	33.39
5.0	10	5.9	41.00
10.0	14.25	7.6	46.67
25	29.1	11.85	59.28
50	53	18.05	65.94
100	117.1	28.95	75.28
250	270.85	36	86.71

ND= Not Detected (below 1st analytical standard)

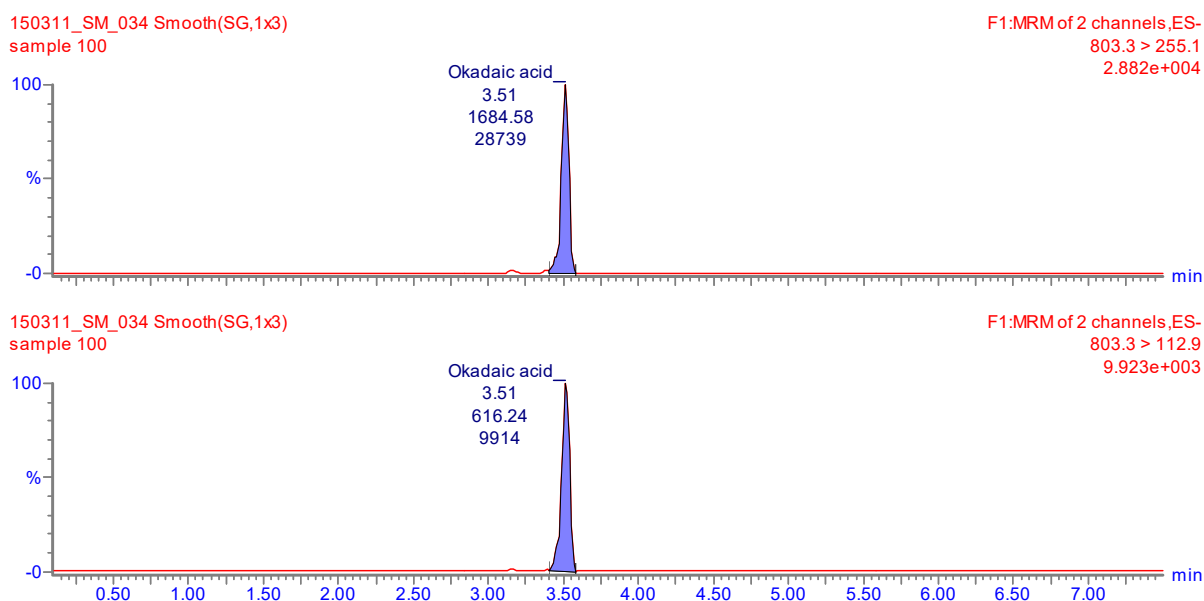


Fig 4: LC-MS/MS Chromatogram of non PL treated 100µg/L OA

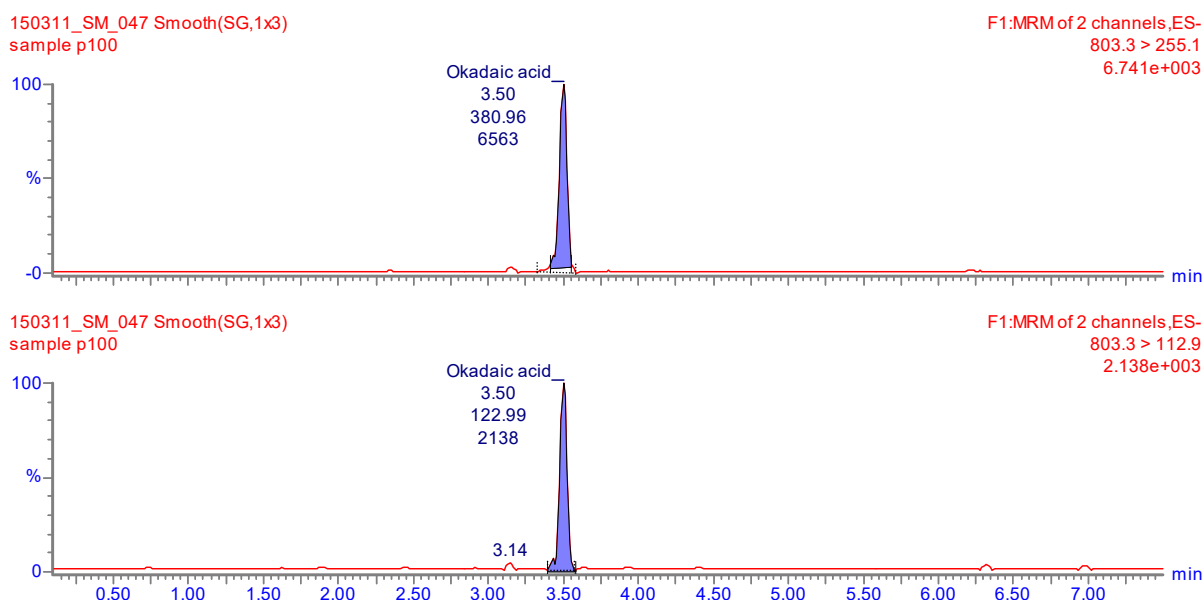


Fig 5: LC-MS/MS Chromatogram of PL treated 100µg/L OA

Discussion

Given the economic losses incurred by the aquaculture industry it is prudent to investigate remedial and preventative processes to mitigate these losses. To this end, the efficacy of PL to reduce the toxicity of OA to *Daphnia pulex* was investigated. The use of PL effectively reduced *Cryptosporidium parvum* by $4 \log^{10}$ orders and subsequent ecotoxicological assessments revealed that PL could be safely used for water decontamination applications, which highlights the importance of conducting tandem environmental testing. In terms of *in situ* water disinfection studies for the fisheries industry, recent reports show that PL can be scaled up for treating contaminated water at wastewater treatment plant level (Garvey and Rowan, 2015; Rowan *et al.*, 2016).

A high UV dose of 12.98 $\mu\text{J}/\text{cm}^2$ was used in this study for investigating efficacy of PL for reducing okadaic acid as this dosage was reported to be effective against recalcitrant bacterial endospores (Garvey and Rowan, 2015), which are more resistant to UV irradiation compared to that of vegetative bacteria and other microorganisms. Selection of this UV dosage was also based on related findings of Farrell et al. (2011), where these researchers reported that irreversible destruction of pathogenic yeast occurred at this upper PL-intensity due to inflicting a multi-hit destructive processes at the cellular and molecular level. This is in marked contrast to conventional UV delivery approaches that rely on causing damage to genetic material where treated microorganisms are often capable of mounting an adaptive protective effect (Rowan *et al.*, 2016). While this study demonstrated potential for reducing the toxicity of algal toxin to freshwater crustacean *Daphnia pulex*, future intensive studies are merited to explore the efficacy of PL technology for treating a broad range of algae and phytoplankton toxins as this would have implications in terms of disinfection potential.

As the test performed is an adaptation of the classical ISO method necessitated by the need for reduced volumes due to the cost of purified OA, correlation between the two test formats was established using the ISO prescribed reference chemical $\text{K}_2\text{Cr}_2\text{O}_7$. As indicated in Figure 3, a Pearson correlation coefficient of 0.9883 was observed with a confidence interval of 0.9344 – 0.9980.

This study demonstrates that the use of PL resulted in more than a 15-fold reduction in the toxicity of OA to *Daphnia pulex* (24hr EC_{50} for untreated OA was 1.68 $\mu\text{g}/\text{L}$ and for PL treated OA was 25.87 $\mu\text{g}/\text{L}$). A concomitant reduction of 86% at 250 $\mu\text{g}/\text{L}$, reducing to 75% and 65% at 100 $\mu\text{g}/\text{L}$ and 50 $\mu\text{g}/\text{L}$ respectively was confirmed by LC-MS/MS analysis.

Although the LC-MS/MS analysis included for DTX-1 and none was observed, further studies are warranted to elucidate the derivatives of PL treated OA. This study demonstrates the novel application of PL to significantly reduce the toxicity of OA to *Daphnia pulex*. Potentially, this non-chemical method could be expanded to treat a broad range of algal toxins.

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